

Air pollution, deprivation and health: Understanding relationships to add value to local air quality management policy and practice in Wales, UK

Brunt H^{1,2*}, Barnes J¹, Jones SJ², Longhurst JWS¹, Scally G³, Hayes E¹

¹ Air Quality Management Resource Centre, University of the West of England, Bristol, UK

² Health Protection Team, Public Health Wales, Cardiff, UK

³ WHO Collaborating Centre for Healthy Urban Environments, University of the West of England, Bristol, UK

* corresponding author

Introduction

Exposure to air pollutants such as nitrogen dioxide (NO₂) and particulate matter (PM₁₀ and PM_{2.5}) is linked with adverse health effects such as heart disease and stroke, respiratory disease and lung cancer.¹ On average, air pollution reduces the life expectancy of every resident in the United Kingdom (UK) by 7-8 months.² The health burden is substantial: 29,000 deaths and 307,000 lost life-years,³ and 23,500 deaths and 277,000 lost life-years,⁴ are attributed annually to PM_{2.5} and NO₂ exposure, respectively in the UK.

While these headlines provide scope and profile to the UK air pollution-and-health problem, they mask important local-level variations in air pollution concentrations, exposures, risks and impacts. Air pollution concentrations (especially NO₂) vary across small geographies because they are influenced by local factors such as traffic, domestic, industrial and agricultural sources, as well as by more-distant sources and meteorological conditions. The differential exposure that results can interact with individual and population-level susceptibilities – ‘intrinsic’ (e.g. age, sex, genetics) and/or ‘acquired’ (e.g. income, education, housing, employment, service access, lifestyle/behaviour-related chronic illnesses)⁵ – to influence health risks and impacts.⁶ Thus, a *triple jeopardy* exists where air pollution, impaired health and deprivation can combine to create increased and disproportionate disease burdens between and within regions.^{7,8}

Given these relationships, regarding local air pollution problems as isolated concerns is a mistake; they should be considered in the broadest possible public health context.⁹ However, this is rarely recognised or realised. In the UK, the Local Air Quality Management (LAQM) regime provides the statutory framework to support collaborative action to assess and reduce local air pollution problems to protect health. Despite these intentions LAQM has historically failed to acknowledge the interaction between wider health determinants and air pollution, their influence on health risks and impacts, and how addressing these together may help solve problems.¹⁰ As a consequence, LAQM and public health policy and practice have remained largely separate realms. This disconnect is detrimental; ill-informed decisions and ineffective or poorly-targeted actions taken based on a limited understanding of such relationships (or worse ignoring them altogether) can compound problems.¹¹

Taking more-informed, targeted and co-ordinated action based on a good understanding of air pollution and health relationships could add value to existing LAQM arrangements and result in greater reductions in air pollution and population-level risks and inequalities. To test this hypothesis, and to inform the future development of LAQM-related policy and practice, this study assesses and quantifies associations between local-level air pollution and relevant health outcomes in the context of deprivation.

Methods

The setting for this ecological cross-sectional study was Wales, a UK principality with a population of 3.1 million people, localised air pollution problems resulting from mixed pollution sources,¹² and high levels of deprivation and health inequalities.¹³ Wales was preferred for this study over other parts of the UK for three reasons. First, all 22 Welsh Local Authorities have equal status; since LAQM policy has been devolved from UK to Welsh Government its implementation through Local Authorities is consistent. Second, in contrast to other parts of the UK, the specialist public health function remains embedded in National Health Service structures. Third, new legislation – the Wellbeing of Future Generations (Wales) Act 2015 – places new statutory responsibilities on public bodies to collaborate to work sustainably to improve the social, economic, environmental and cultural well-being.¹⁴ Unique to Wales, this new legislative framework offers opportunities to connect and enhance air quality management and public health policy and practice that do not exist elsewhere.

Given this study's emphasis on assessing local-level associations, data corresponded to Wales' smallest (highest resolution) administrative geographies called Lower-layer Super Output Areas (LSOA). These approximate 'neighbourhood' areas, of which there are 1,909 in Wales, have an average 1,600 residents (range: 1,000 to 3,000) and 650 households (range: 400 to 1,200) and take into account proximity and social homogeneity.

The challenge of assessing air pollution exposure was overcome by using modelled data as these more-accurately reflect area-level exposure than the relatively crude use of measurements from monitors nearest to populations.¹⁵ Modelled annual mean ambient NO₂, PM₁₀ and PM_{2.5} concentration data at 1x1km-grid resolution were averaged for the 2011-2013 study period. These data were obtained from the UK Government's Pollution Climate Mapping model. This model generates validated annual estimates of area-level pollutant concentrations (based on 2011 as a baseline year, and projected annually) through land-use regression and kernel-based air dispersion modelling of known emission sources that takes account of meteorological conditions.¹⁶ Prior to their use in this study, air pollution data were converted to population-weighted LSOA exposure estimates using standard methods.¹⁷

Records of all mortality (not only hospital deaths) for years 2011-2013 were obtained from Office for National Statistics through the NHS Wales Information Service. Numbers of deaths for health outcomes of interest in the context of air pollution exposure were identified by using appropriate International Classification of Diseases (ICD, version 10) codes. These were: all-cause non-accidental mortality (excluding injuries and external causes) (ICD-10: A00-R99), cardiovascular diseases (ICD-10: I00-I99), cerebrovascular diseases (ICD-10: I60-I69) and respiratory diseases (ICD-10: J00-J99). Mortality data for chronic liver disease (ICD10: K70, K73, K74) were also obtained to act as a 'control' outcome¹⁸ since this outcome is known to be influenced by deprivation-related risk factors¹⁹ but not by air pollution (noting emerging evidence from animal studies²⁰). Hospital admissions data – records of all inpatient and day case activity undertaken in NHS Wales (morbidity) – were obtained from the Patient Episode Database for Wales for the same health outcomes, and for the same three-year study period. All 'health' data were then stratified by five-year age bands and linked to study-period-averaged mid-year LSOA population estimates obtained from the Office for National Statistics.²¹

Deprivation data were obtained from Welsh Government's Welsh Index of Multiple Deprivation (WIMD).²² The WIMD assigns each LSOA in Wales a summary deprivation score derived from a weighted combination of data from eight domains: income (23.5%); employment (23.5%); health (14%); education (14%); geographical access to services (10%); community safety (5%); housing (5%) and physical

environment (5%). Each of these domains includes several indicators of deprivation e.g. income-deprivation is a composite measure reflecting the proportion of all residents of an LSOA with income below a defined level; it is calculated from LSOA numbers of income-related benefit claimants, tax credit recipients and supported asylum seekers. For this study, it was inappropriate to use the LSOA-level summary WIMD scores since their composition had been influenced by health and air pollution data. To avoid introducing bias from 'double-counting' these component data, income-deprivation domain data were used as an indicator of multiple deprivation.^{18,23-25}

There were two phases of data analysis:

1. Linking and describing LSOA data

Each LSOA was assigned one of five income-deprivation status classifications. Quintiles were derived by ranking income-deprivation composite scores for all LSOAs and dividing the data into five roughly equal parts (each with 381 or 382 LSOAs and a population of approximately 600,000 people).

LSOAs were also assigned an air pollution status classification of being a 'low', 'moderately' or 'high' polluted area. Cut-off points for tertiles were determined by ordering the distribution of LSOA air pollution concentrations (for each pollutant) and dividing the data falling between the 5th and 95th percentile values into three equal parts. LSOAs with data values below the 5th percentile (n=40 for NO₂, n=97 for PM₁₀, n=90 for PM_{2.5}) or above the 95th percentile (n=94 for NO₂, n=90 for PM₁₀, n=77 for PM_{2.5}) were assigned either 'low' or 'high' polluted area status, as appropriate.

Area-level air pollution and income-deprivation status data, and mortality and hospital admissions data, were linked by LSOA using Microsoft Excel and ArcGIS 10.2.2 software. Linked data were subsequently aggregated based on deprivation and area-level NO₂, PM₁₀ and PM_{2.5} air pollution status. Using mortality and hospital admissions numerator data and mid-year population denominator estimates, European Age-Standardised Rates (EASR) with 95% confidence intervals were calculated for each health outcome.^{26,27} Through this direct method of standardisation, age-adjusted rates were derived by applying crude mortality or hospital admissions rates for each health outcome of interest (calculated post data aggregation) to a single reference population, in this case the European standard population. The result was standardised rates, adjusted for differences in the age structure of the population, which facilitated comparisons over time and place.

2. Assessing associations

Air pollution-health associations and, separately, *deprivation-health* associations, were assessed using rate ratios (RR) with 95% confidence intervals.^{28,29} RRs compared rates of health outcomes in 'high' polluted or 'most' income-deprived areas with those in reference 'low' polluted or 'least' deprived areas. The *air pollution-deprivation-health* association assessment – which considered air pollution and deprivation interactions and their combined association with health outcomes – adopted the same method to compare the rates in reference 'low' polluted and 'least' deprived areas with elsewhere.

Results

Results are presented below for each data analysis phase.

1. Linking and describing LSOA data

Across all LSOAs, the annual mean NO₂ concentration was 17.7 µg/m³ (5th percentile: 6.6 µg/m³; 95th percentile: 36.7 µg/m³), for PM₁₀ it was 14.1 µg/m³ (5th percentile: 11.4 µg/m³; 95th percentile: 17.3 µg/m³), and for PM_{2.5} it was 9.5 µg/m³ (5th percentile: 7.5 µg/m³; 95th percentile: 11.8 µg/m³). Local-level NO₂ concentrations varied substantially (Fig. 1). Concentrations were greatest over town and city agglomerations in south-east Wales, along main traffic routes such as the M4 motorway running between the south-east and mid-south Wales, and in heavily-industrialised areas like the busy sea port where several oil refineries are located in south-west Wales. Local-level PM₁₀ and PM_{2.5} concentrations also varied, but not to the same extent (not shown).

Income-deprivation status also varied at the local level (Fig. 1). The income-deprivation composite scores that lie under the LSOA classifications revealed that the proportion of people living in income deprivation ranged from 5% in some LSOAs to 31% in others (all-Wales average: 16%). Only 12% of 'most' deprived areas could be described as being rural compared with 27% of 'least' deprived areas; 'next least' deprived areas were least urbanised.

In the context of deprivation, a marked 'u'-shaped, non-linear NO₂-value distribution pattern was observed across quintiles; average concentrations were highest in 'most' deprived areas and next highest in 'least' deprived areas (Fig. 2). The patterns for PM₁₀ and PM_{2.5} were similar to that of NO₂ but less pronounced.

In the context of health, the average annual all-cause non-accidental death count in Wales was 30,035 (Wales EASR=100.5 per 10,000; 95%CI: 99.4-101.7). Cardiovascular disease accounted for 31% of these (EASR=30.9 per 10,000; 95%CI: 30.3-31.6), respiratory disease 16% (EASR=15.7 per 10,000; 95%CI: 15.3-16.2), cerebrovascular disease 8% (EASR=7.6 per 10,000; 95%CI: 7.3-8.0) and chronic liver disease 1% (EASR=1.4 per 10,000; 95%CI: 1.3-1.6). For each death, there were 3.3, 9.7, 2.3, and 2.2 times as many hospital admissions for the same diseases, respectively.

2. Assessing associations

The *air pollution-health* association analysis revealed that rates of only all-cause non-accidental mortality (RR=1.27; 95% CI: 1.10-1.45) and respiratory disease mortality (RR=1.43; 95% CI: 1.03-1.96) increased as NO₂ air pollution worsened and were significantly higher in 'high' polluted areas compared with 'low' polluted areas (Table 1). Similar associations were observed for PM_{2.5} (all-cause non-accidental mortality: RR = 1.15; 95% CI: 1.10-1.20; respiratory disease mortality: RR = 1.37; 95% CI: 1.22-1.52) and PM₁₀ (all-cause non-accidental mortality: RR = 1.14; 95% CI: 1.08-1.20; respiratory disease mortality: RR = 1.31; 95% CI: 1.15-1.50; respiratory disease morbidity: RR = 1.17; 95% CI: 1.12-1.21).

The *deprivation-health* association analysis showed that income-deprivation status was positively and significantly associated with all health outcomes, especially chronic liver disease mortality and morbidity (Fig. 3). *Deprivation-health* associations were stronger than *air pollution-health* associations. With the exception of the 'control' chronic liver disease outcomes, income-deprivation status was most strongly

associated with respiratory disease mortality (RR=1.97; 95%CI: 1.79-2.17) and morbidity (RR=2.05; 95%CI: 1.98-2.11).

As for *air pollution-deprivation-health* association assessment, when considered in the context of air pollution, positive associations between deprivation status and health persisted (Table 2). All health endpoints were positively associated with income-deprivation status; rates were higher in 'most' deprived/'low' polluted areas than in reference 'least' deprived/'low' polluted areas, regardless of air pollution status. Chronic liver disease outcomes continued to be most strongly associated with deprivation status, followed by respiratory disease outcomes. Simultaneously considering income-deprivation and air pollution status (for PM₁₀ and PM_{2.5}, but not NO₂) strengthened the associations observed previously in the *deprivation-health*-only analysis for only all-cause non-accidental and respiratory disease mortality.

In 'least' deprived areas, rates of all-cause non-accidental and respiratory disease mortality increased as NO₂, PM₁₀ and PM_{2.5} air pollution status worsened, but associations were non-significant. In 'most' deprived areas, strong positive associations were observed between all air pollutants and all-cause non-accidental and respiratory disease mortality (Table 1). In these latter areas, for these health outcomes, air pollution increased the significance of *deprivation-health* associations. To illustrate this with an example drawn from the data: for PM₁₀, the respiratory mortality rate was a significant 2.05 times higher in 'low' polluted/'most' deprived areas (RR: 2.05; 95%CI: 1.73-2.41) compared with reference 'low' polluted/'least' deprived areas. Across 'most' deprived areas, as air pollution status worsened associations were strengthened, becoming 2.21 times higher (RR: 2.21; 95%CI: 1.92-2.53) in 'moderately' polluted areas and 2.38 times higher (RR: 2.38; 95%CI: 1.89-2.95) in 'high' polluted areas.

It should be noted that significant negative associations were found in 'most' income-deprived areas between: NO₂ and respiratory disease morbidity; PM₁₀ and cardiovascular disease mortality, cerebrovascular and chronic liver disease morbidity; and PM_{2.5} and cardiovascular disease mortality.

Discussion

Main findings

Air pollution concentrations, especially NO₂, showed LSOA-level variation.. Average air pollution concentrations were relatively high in both 'most' and 'least' deprived areas, but were highest in the former. Substantial local-level deprivation-related health inequalities were observed; the magnitude of *deprivation-health* associations was greater than *air pollution-health* associations. That said, not accounting for deprivation status, each pollutant was positively and significantly associated with all-cause non-accidental and respiratory disease mortality, and PM₁₀ with respiratory disease morbidity too. When considered simultaneously, the interaction between air pollution and deprivation status modified and amplified associations with all-cause non-accidental and respiratory disease mortality endpoints, especially in 'most' deprived areas where Wales' most-vulnerable populations live. While action is needed to reduce air pollution concentrations and associated risks everywhere, for these health outcomes in these areas, lowering air pollution and deprivation status to that of 'low' polluted and 'least' deprived areas could achieve a substantial additional health gain.

What is already known?

The evidence for a socio-economic gradient in health is well-established. The average seven-year life-expectancy difference between 'most' and 'least' deprived areas in Wales¹³ and the UK³⁰ is mostly

attributed to multiple deprivation risk factors, especially lifestyle behaviours and choices.³¹ This study corroborated findings that *deprivation-health* associations are stronger than *air pollution-health* associations.¹⁸ However, as also found here, air pollution is a known environmental health determinant that adds to already-strong *deprivation-health* associations.³² This is supported by unequivocal evidence of independent, likely causal relationships between air pollution exposure and cardio-pulmonary and other health impacts.¹

Several studies have assessed air pollution and deprivation associations. In the US, Canada and New Zealand, higher air pollution levels have been reported in socioeconomically-disadvantaged compared with less-deprived communities.³³⁻³⁸ However, the situation in Europe appears to be less straightforward; findings from studies across Europe have generated mixed results.³⁹⁻⁴⁵ In the UK, Walker *et al.* previously reported findings that are consistent with those of this study, that both 'most' and 'least' deprived areas were disproportionately affected by high NO₂ concentrations.⁴² A number of possible explanations for these inconclusive research findings have been offered; all relating to characteristics of urbanised areas. For example, a study comparing local, regional and national-level associations between air pollution and socioeconomic factors in England and the Netherlands suggested that more-deprived areas are often in close proximity to mixed/high-traffic roads.²³ A study exploring the same relationships (at Local Authority level) in England and Wales explained that areas of mixed deprivation are often adjacently-located in urban areas,⁴³ which may be the result of city gentrification and land-use planning decisions.²³ Lastly, although beyond the UK context, a study that examined the environmental inequity of traffic-related air pollution in Toronto, Canada, proposed those living in 'least deprived' urban areas tolerate more pollution in lieu of living, social and employment benefits.⁴⁶

A number of studies have also explored air pollution and deprivation associations in the context of vulnerable people affected. Environmental justice analyses of air quality in the UK have found that children are disproportionately exposed (and are more vulnerable) to higher levels of air pollution.^{44,47} This present study found that 'most deprived' areas contained the highest proportion of children aged <15 years in any deprivation quintile (24%; estimated population: 122,458). Additionally, Laxen *et al.*,⁴⁸ unlike Fecht *et al.*,²³ found that older people were disproportionately exposed to air pollution in the city of Bristol, UK. This study identified that 'least' deprived areas had high air pollution concentrations and contained the highest proportion of older people aged 75+ years in any deprivation quintile (23%; estimated population: 57,332). Despite having a high proportion of older people, it should be noted that less-deprived populations are generally healthier and so are likely to be less susceptible to the effects of air pollution.⁴⁹

A relatively small number of studies have extended the research boundaries beyond air pollution and deprivation to also consider health impacts. These have generally found that area-level air pollution status is associated with significant increases in all-cause non-accidental, cardiovascular and respiratory disease mortality risk in high-deprivation compared with less-deprived areas.^{18,25,50-56} Weak relationships of air pollution and deprivation with morbidity (respiratory hospital admissions) have been reported.^{51,57} While it is often difficult to compare studies because of inconsistencies in research approaches and methods, in the main, interactions between air pollution and socio-economic factors have been found to modify and compound the health effects associated with each variable individually.^{53,58,59}

To summarise, a balanced review of European evidence suggests the general pattern is: that deprived people, while not always exposed to higher levels of air pollution, are more likely to suffer greater harm as a consequence of their exposure since they are more vulnerable to its effects.⁴⁵

What this study adds

Prior to this study, little was known about the relationships between air pollution, deprivation and health in Wales, especially at the local level. Findings raise important issues that should now inform debates around the future development of LAQM and public health policy and practice in Wales and beyond. The significant, and complex, interactions between these variables justify air pollution problems and solutions not being considered in isolation. Rather, air pollution problems should be regarded as a local public health priority that is inextricably linked with other behavioural, societal and environmental determinants of health. For example, people living in poorly-designed communities where sustainability and active travel is not promoted may be overly-dependent on the use of cars; this may lead to physical inactivity that, when coupled with other more-likely behaviours in most deprived areas such as high alcohol consumption, poor diet and smoking, increases the risk of poor health outcomes like cardio-respiratory diseases, obesity, diabetes, cancer and mental ill-health. This poor health status makes people susceptible to the effects of air pollution (which is highest in most deprived areas) and exacerbates problems.

It remains important to reduce air pollution exposures and associated risks for all. However, this study shows that greater public health gains (maximised risk reduction and minimised inequalities) can result not from simply reducing air pollution and exposure potential, but by doing so alongside efforts to tackle broader health determinants in areas where health needs are greatest. Typically, this is in 'most' deprived and polluted areas where, as this study found, rates of cardio-respiratory mortality could be up to 2.4 times lower if air pollution and income-deprivation status were reduced to those of 'low' polluted and 'least' deprived areas.

Considering local problems and solutions in the broadest possible public health context must be the priority. Air pollution mitigation and public health intervention should be evidence-informed, targeted and co-ordinated. Public health agencies and professionals have a valuable contribution to make to this work. Specifically, in the context of LAQM policy and practice, greater public health integration and engagement in LAQM can add value through:

- Risk assessment – data sharing, linkage, analysis and interpretation can improve population risk assessments. Results will be more comprehensive, accurate and meaningful in terms of scoping problems, defining at-risk populations, and understanding relationships, causes and solutions. Appreciating this 'big picture' can, in turn, inform targeted risk communications and interventions.
- Management – collaborating to achieve air quality and public health co-benefits through active travel interventions (encouraging walking and cycling over vehicle use). Also, implementing strategies to improve individual and community baseline-health status and thus reduce susceptibility (e.g. promoting nutrition, smoking cessation, service access), educating people to change behaviours and advocating for the design of sustainable and healthy communities, and the separation of people and pollution sources, by working with local Planning Authorities to inform decision-making.⁶⁰
- Intervention and policy development – facilitating intervention evaluations (that consider air pollution reduction *and* health impacts together) to determine what works in real-world situations. Also, advocating for, and providing authority, leadership and autonomy to bring about evidence-based change through stronger policy connection and change.

Limitations

A limitation of ecological studies generally is assuming that hazards, risks and outcomes are spread evenly across defined areas and populations. Since this study assessed air pollution, deprivation and health associations by using aggregated area-level data, the risk of the ‘ecological fallacy’⁶¹ – making inferences about individuals from area/population-level analyses – was minimised. Assumptions around homogeneity, and separate air pollutant analyses that did not account for pollutant synergies,⁶² may have actually underestimated the strength of the local-level associations found. It is also important to note that, as a cross-sectional study, causal inference cannot be drawn from the findings.

This study’s emphasis on exploring ‘local’ associations meant that, as with all small-area analyses, there were limitations of using high-resolution data. Limitations include: numbers of health events over short periods may be small and give rise to health burden estimates affected by chance and random variation, measures of exposure are approximate, boundaries and populations may change over short time-frames, and allocating events to areas is difficult. In some instances the relatively small numbers of health events gave rise to some results that were not expected such as the significant negative associations between area-level PM status and cardiovascular disease mortality. However, in the main, selecting LSOAs as the geographical study unit, and linking all data at this level, made this study more robust. Data were aggregated based on matched area-level characteristics (air pollution and income-deprivation status) to avoid problems associated with small numbers, data for three years were used to smooth annual variations, and LSOA boundaries were consistent over the study period. Also, despite varying in size geographically, the limited population variation of LSOAs facilitated comparisons across small areas, compared with geographic units used in ecological analyses historically such as electoral ‘wards’.

Modelled air pollution data were preferred over measured data from discrete monitoring points (where distances, and probability of pollution variation, between receptor and the nearest monitor can be significant) as they more accurately reflect area concentrations and population exposures.^{15,55} The modelled data used in this study are validated annually against measured data from air pollution monitoring stations. While this increases confidence in area-level modelled air pollution data, it should be noted that such area-level estimates may not correspond exactly with actual personal exposure which is influenced by an individual’s mobility, time spent indoors and outdoors, and activity patterns, levels and types. Given that this ecological study was concerned with area-level relationships only, it was inappropriate to attempt to estimate the exposure of individuals; this avoided bias from exposure misclassification.

Only one measure of deprivation – income deprivation – was used as an indicator for area-level multiple deprivation status. Doing so avoided ‘double counting’ the health and air pollution components that would have occurred had the summary WIMD score been used and therefore minimised the possibility of delivering skewed results.^{7,18,23}

It was not possible to account for all confounding factors. Smoking, for example, is a key risk factor for the health outcomes of interest in this study, but only Local Authority-level smoking prevalence data were available which were based on self-reported survey responses from a sample of the Welsh population. Using data such as these (not specific to small areas), would have yielded little bias-reduction since large-area summary data are often grossly inadequate to ensure effective control.⁶³ Attempting to disaggregate these data would have introduced greater uncertainty and bias. Other studies that have explored air pollution, deprivation and health associations have confirmed that

adjusting for smoking behaviour where available at larger geographies does not significantly attenuate results.^{50,51,57} The use of chronic liver disease ‘control’ outcomes did however inform the interpretation of results.

Given that some important data were not available by LSOA, a simple approach to investigating local-level relationships was selected over complex regression techniques. This approach is intended to be understandable and meaningful to policy-makers and the public. However, it may be perceived by some to be overly-simplistic. In response, it is argued that, although a small-area study, the denominator population is over 3 million people – a sample size that reduces the effect of confounders. Additionally, population variations are accounted for to a large extent through age-standardisation.

Conclusion

Air pollution concentrations were highest in ‘most’ deprived areas. Separately, *deprivation-health* associations were stronger than *air pollution-health* associations. Considered simultaneously, air pollution added to *deprivation-health* associations for some health outcomes. Interactions between air pollution and deprivation modified and amplified associations with all-cause non-accidental and respiratory disease mortality, especially in ‘most’ deprived areas where Wales’ most-vulnerable populations live.

Air pollution, deprivation and health are inextricably linked. There is merit in implementing measures to reduce air pollution concentrations and exposures for all, just as there is in tackling deprivation-related risk factors. However, if local air pollution problems and solutions are considered in the context of wider health determinants, and air pollution mitigation and public health interventions are aligned and targeted in areas where health needs are highest, greater health gains (reduced health risks and inequalities) can be achieved. To facilitate this, achieving greater public health integration and engagement in LAQM policy and practice must be prioritised in Wales and beyond.

Acknowledgements

The authors would like to thank Dr Marion Lyons (Public Health Wales), Dr Tim Chatterton and Prof Jimi Irwin (both University of the West of England) for their support, advice and guidance. This research was part-funded by Public Health Wales.

References

- 1 WHO. *Review of evidence on health aspects of air pollution-REVIHAAP*. 2013. Copenhagen: WHO. http://www.euro.who.int/_data/assets/pdf_file/0004/193108/REVIHAAP-Final-technical-report-final-version.pdf?ua=1 (12 May 2016, date last accessed).
- 2 Defra. *The air quality strategy for England, Scotland, Wales and Northern Ireland (vol 1)*. 2007. London: Defra. https://www.gov.uk/government/uploads/system/uploads/attachment_data/file/69336/pb12654-air-quality-strategy-vol1-070712.pdf (12 May 2016, date last accessed).
- 3 Gowers AM, Miller BG, Stedman JR. *Estimating local mortality burdens associated with particulate air pollution*. 2014. London: Public Health England. https://www.gov.uk/government/uploads/system/uploads/attachment_data/file/332854/PHE_CRCE_010.pdf (12 May 2016, date last accessed).
- 4 Defra. *Draft plans to improve air quality in the UK: tackling nitrogen dioxide in our towns and cities. UK overview document*. 2015. London: Defra. https://consult.defra.gov.uk/airquality/draft-aq-plans/supporting_documents/Draft%20plans%20to%20improve%20air%20quality%20in%20the%20UK%20%20Overview%20document%20September%202015%20final%20version%20folder.pdf (12 May 2016, date last accessed).
- 5 Fann N, Roman HA, Fulcher CM *et al*. Maximising health benefits and minimizing inequality: incorporating local-scale data in the design and evaluation of air quality policies. *Risk Analysis* 2011; DOI:10.1111/j.1539-6924.2011.01629.x.
- 6 Lipfert FW. Air pollution and poverty: does the sword cut both ways? *Journal of Epidemiology and Community Health* 2004; 58: 2-3.
- 7 Goodman A, Wilkinson P, Stafford M, Tonne C. Characterising socioeconomic inequalities in exposure to air pollution: a comparison of socio-economic markers and scales of measurement. *Health & Place* 2011; 17:767-774.
- 8 Jerrett M, Burnett R, Kanaroglou P, *et al*. A GIS-environmental justice analysis of particulate air pollution in Hamilton, Canada. *Environment and Planning A* 2001; 33: 955-973.
- 9 O'Neill MS, Jerrett M, Kawachi *et al*. Health, wealth and air pollution: advancing theory and methods. *Environmental Health Perspectives* 2003; 111(16): 1861-1870.
- 10 Brunt H, Barnes J, Longhurst JWS *et al*. Local Air Quality Management policy and practice in the UK: the case for greater Public Health integration and engagement. *J Env Science and Policy* 2016; 58: 52-60.
- 11 Bowen W. An analytical review of environmental justice research: what do we really know? *Environmental Management* 2002; 29(1): 3-15.
- 12 Defra. *Air Quality Management Area summary data*. 2016. London: Defra. <https://uk-air.defra.gov.uk/aqma/summary> (12 May 2016, date last accessed).

- 13 Welsh Government. *Chief Medical Officer Annual Report 2014-15: Happier, Healthier, Fairer*. 2015. Cardiff: Welsh Government. <http://gov.wales/docs/phhs/publications/160323cmoreporten.pdf> (12 May 2016, date last accessed).
- 14 Welsh Government. *Wellbeing of Future Generations (Wales) Act 2015: the essentials*. 2015. Cardiff: Welsh Government. <http://gov.wales/docs/dsjlg/publications/150623-guide-to-the-fg-act-en.pdf> (12 May 2016, date last accessed).
- 15 Krewski D, Jerrett M, Burnett RT *et al*. Extended follow-up and spatial analysis of the American Cancer Society study linking particulate air pollution and mortality. *Res Rep Health Eff Inst* 2009; 140: 5-114.
- 16 Stedman J, Bush T, Vincent KJ *et al*. *UK air quality modelling for annual reporting 2002 on ambient air quality assessment under Council Directives 96/62/EC and 1999/30/EC*. 2003. Abingdon, UK: AEA Technology. https://uk-air.defra.gov.uk/assets/documents/reports/cat05/0402061100_dd12002mapsrep1-2.pdf (12 May 2016, date last accessed).
- 17 Welsh Government (2014). *Welsh Index of Multiple Deprivation 2014 (WIMD 2014) Technical Report*. 2014. Cardiff, UK: Welsh Government. <http://gov.wales/statistics-and-research/welsh-index-multiple-deprivation/technical-information/?lang=en> (18 May 2016, date last accessed).
- 18 Richardson EA, Pearce J, Tunstall H *et al*. Particulate air pollution and health inequalities: a Europe-wide ecological analysis. *International Journal of Health Geographics* 2013; 12: 34-43.
- 19 Major JM, Sargent JD, Graubard BI *et al*. Local geographic variation in chronic liver disease and hepatocellular carcinoma: contributions of socioeconomic deprivation, alcohol retail outlets and lifestyle. *Annals of Epidemiology* 2014; 24: 104-110.
- 20 Kim JW, Park S, Lim CW *et al*. The role of air pollutants in initiating liver disease. *Toxicological Research* 2014; 30(2): 65-70.
- 21 Office for National Statistics. *Lower Super Output Areas Population Estimates for Wales*. 2015. Newport, UK: ONS. <http://www.ons.gov.uk/peoplepopulationandcommunity/populationandmigration/populationestimates/datasets/lowersuperoutputareamidyearpopulationestimates> (12 May 2016, date last accessed).
- 22 Welsh Government. *Welsh Index of Multiple Deprivation*. 2016. Cardiff, UK: Welsh Government. <http://wimd.wales.gov.uk/> (12 May 2016, date last accessed).
- 23 Fecht D, Fischer P, Fortunato L, Hoek G *et al*. Associations between air pollution and socioeconomic characteristics, ethnicity and age profile in neighbourhoods in England and the Netherlands. *Environmental Pollution* 2015; 198: 201-210.
- 24 Kruize H, Driessen PPJ. Environmental equity and the role of public policy: experiences in the Rijnmond region. *Environmental Management* 2007; 40: 578-595.
- 25 Næss O, Piro FN, Nafstad P *et al*. Air pollution, social deprivation and mortality: a multilevel cohort study. *Epidemiology* 2007; 18(6): 686-694.

- 26 Armitage P, Berry G. *Statistical methods in medical research* (3rd edn). 1994. Oxford: Blackwell.
- 27 Breslow NE, Day NE. *Statistical methods in cancer research, volume II: The design and analysis of cohort studies*. 1987. Lyon: International Agency for Research on Cancer, World Health Organisation.
- 28 Newcombe RG. MOVER-R confidence intervals for ratios and products of two independently estimated quantities. *Statistical Methods in Medical Research* 2013; 0(0): 1-5.
- 29 Newcombe RG, Bender R. Implementing GRADE: calculating the risk difference from the baseline risk and the relative risk. *Evid Based Med* 2013; 0(0):1-4.
- 30 Marmot M. *Fair Society, Healthy Lives: the Marmot Review*. 2010. London: UCL Inst of Health Equity. <https://www.instituteofhealthequity.org/projects/fair-society-healthy-lives-the-marmot-review> (12 May 2016, date last accessed).
- 31 Frohlich KL, Abel T. Environmental justice and health practices: understanding how health inequalities arise at the local level. *Sociology of Health and Illness* 2014; 36(2): 199-212.
- 32 Lim SS, Vos T, Flaxman AD *et al*. A comparative risk assessment of burden of disease and injury attributable to 67 risk factors and risk factor clusters in 21 regions, 2990-2010: a systematic analysis for the Global Burden of Disease study 2010. *The Lancet* 2012; 380: 2224-2260.
- 33 Brochu PJ, Yanosky JD, Paciorek *et al*. Particulate air pollution and socioeconomic position in rural and urban areas of northeastern United States. *American Journal of Public Health* 2011; 101(1): 224-230.
- 34 Crouse DL, Ross NA, Goldberg MS. Double burden of deprivation and high concentrations of ambient air pollution at the neighbourhood scale in Montreal, Canada. *Soc Sci Med* 2009; 69(6): 971-981.
- 35 Pearce J, Kingham S. Environmental inequalities in New Zealand: a national study of air pollution and environmental justice. *Geoforum* 2008; 39: 980-993.
- 36 Yanosky JD, Schwatz J, Suh HH. Associations between measures of socioeconomic position and chronic nitrogen dioxide exposure in Worcester, Massachusetts. *Journal of Toxicology and Environmental Health, Part A* 2008; 71: 1593-1602.
- 37 Pearce J, Kingham S, Zawar-Reza P. Every breath you take? Environmental justice and air pollution in Christchurch, New Zealand. *Environment and Planning A* 2006; 38: 919-938.
- 38 Su JG, Larson T, Gould T *et al*. Transboundary air pollution and environmental justice: Vancouver and Seattle compared. *GeoJournal* 2009; 75: 595-608.
- 39 Fernandez-Somoano A, Hoek G, Tardon A. Relationships between area-level socioeconomic characteristics and outdoor NO₂ concentrations in rural and urban areas of northern Spain. *BMC Public Health* 2013; 13: 71-81.
- 40 Havard S, Deguen S, Zmirou-Navier D *et al*. Traffic-related air pollution and socioeconomic status: a spatial autocorrelation study to assess environmental equity on a small-area scale. *Epidemiology* 2009; 20: 223-230.
- 41 Briggs D, Fecht D, De Hoogh K. Environmental inequity in England: small area associations between socio-economic status and environmental pollution. *Soc Sci and Med* 2008; 67: 1612-1629.

- 42 Walker G, Fairburn J, Smith G, Mitchell G. *Air quality and social deprivation in the UK: an environmental inequalities analysis*. 2006. Bristol: Environment Agency.
- 43 McLeod H, Langford IH, Jones AP *et al*. The relationships between socioeconomic indicators and air pollution in England and Wales: implications for environmental justice. *Reg Env Change* 2000; 1(2): 18-26.
- 44 Mitchell G, Dorling D. An environmental justice analysis of British air quality. *Environment and Planning A* 2003; 35: 909-929.
- 45 Deguen S, Zmirou-Navier D. Social inequalities resulting from health risks related to ambient air quality – a European review. *European Journal of Public Health* 2010; 20(1): 27-35.
- 46 Buzzelli M, Jerrett M. Geographies of susceptibility and exposure in the city: environmental inequity of traffic-related air pollution in Toronto. *Can J Reg Sci* 2007; 30: 195-210.
- 47 Barnes JH, Chatterton T. An environmental justice analysis of exposure to traffic-related pollutants in England and Wales. *Air Pollution XXVI, proceedings of 24th international conference on modelling, monitoring and management of air pollution*. Crete, Greece: 20-22 June 2016.
- 48 Laxen D, Beattie C, Dickie I. *Health impacts of air pollution in Bristol*. 2014. Bristol: Air Quality Consultants.
- 49 Makri A, Stilianakis NI. Vulnerability to air pollution health effects. *Int J Hyg Environmental-Health* 2008; 211: 326-336.
- 50 Richardson EA, Pearce J, Kingham S. Is particulate air pollution associated with health and health inequalities in New Zealand? *Health and Place* 2011; 17: 1137-1143.
- 51 Pearce JR, Richardson EA, Mitchell RJ, Shortt NK. Environmental justice and health: a study of multiple environmental deprivation and geographical inequalities in health in New Zealand. *Social Science and Medicine* 2011; 73: 410-420.
- 52 Pearce JR, Richardson EA *et al*. Environmental justice and health: the implications of the socio-spatial distribution of multiple environmental deprivation for health inequalities in the UK. *Trans Inst Br Geogr* 2010; 35: 522-539.
- 53 Forastiere F, Stafoggia M, Tasco C *et al*. Socioeconomic status, particulate air pollution, and daily mortality: differential exposure or differential susceptibility. *Am J of Industrial Medicine* 2007; 50: 208-216.
- 54 Finkelstein MM, Jerrett M, Sears MR. Environmental inequality and circulatory disease mortality gradients. *Journal of Epidemiology and Community Health* 2005; 59: 481-487.
- 55 Jerrett M, Buzzelli M, Burnett RT, DeLuca PF. Particulate air pollution, social confounders, and mortality in small areas of an industrial city. *Social Medicine* 2005; 60: 2845-2863.
- 56 Villeneuve PJ, Burnett RT, Shi Y *et al*. A time-series study of air pollution, socioeconomic status and mortality in Vancouver, Canada. *J Exposure Analysis and Environmental Epidemiology* 2003; 13: 427-435.

- 57 Wheeler BW, Ben-Shlomo Y. Environmental equity, air quality, socioeconomic status and respiratory health: a linkage analysis of routine data from the Health Survey for England. *Journal of Epidemiology and Community Health* 2005; 59: 948-954.
- 58 Zeka A, Zanobetti A, Schwartz J. Individual level modifiers and the effects of particulate matter on daily mortality. *American Journal of Epidemiology* 2006; 163(9): 849-859.
- 59 Jerrett M, Burnett RT, Brook J *et al.* Do socioeconomic characteristics modify the short term association between air pollution and mortality? Evidence from a zonal time series in Hamilton, Canada. *Journal of Epidemiology and Community Health* 2004; 58: 31-40.
- 60 Giles LV, Barn P, Kunzli N *et al.* From good intentions to proven interventions: effectiveness of actions to reduce the health impacts of air pollution. *Environmental Health Perspectives* 2011; 119(1): 29-36.
- 61 Hart J. On ecological studies: a short communication. *Dose Response* 2011; 9(4): 497-501.
- 62 Walton H, Dajnak D, Beevers S *et al.* Understanding the health impacts of air pollution in London. 2015. London: Kings College London.
<https://www.kcl.ac.uk/lsm/research/divisions/aes/research/ERG/research-projects/HIAinLondonKingsReport14072015final.pdf> (12 May 2016, date last accessed).
- 63 Greenland S, Robins J. Ecological studies – biases, misconceptions and counterexamples. *American Journal of Epidemiology* 1994; 139(8): 747-760.